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Removal of Chloroform from Aqueous Solution by the Adsorption onto Zero Valent Iron Supported on Zeolite as an Efficient Adsorbent

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Abstract

Chloroform is one of the Volatile Organic Compounds (VOCs). Chloroform is the most commonly exist compound of the tri-halo methane (THM); THMs are Halogen-substituted single-carbon compounds with the general formula CHX_3 . The adsorption of chloroform (CHCl_3) from aqueous solution was investigated in series of batch experiment. Two type of adsorbents of zero valent iron were prepared as NZVI and NZVI/ZSM via the reduction method by using sodium boro- hydride as reducing agent. The prepared adsorbent were characterized by the field emission scanning electron microscopy (FESEM), Fourier transform infrared spectroscopy (FTIR), Brunner –Emmett –Teller (BET), and X-ray powder diffraction (XRD). The BET surface areas of the adsorbents were 254.33, 41, and 116.52 m^2/g for ZSM, NZVI, and NZVI/ZSM, respectively. The adsorption isotherm studies were conducted at various initial concentrations of CHCl_3 (10-40) mg/L. Langmuir and Freundlich adsorption isotherm models were employed to fit the adsorption characteristic of the adsorbent. The removal efficiency of CHCl_3 via the NZVI/ZSM from aqueous solution was found to be 19.92 mg/g, which conducted from the Langmuir slope. In addition the adsorption reached equilibrium within 90 min at optimum pH value of 9. It was found that Freundlich isotherm characterized the adsorption process better than that of Langmuir for NZVI/ZSM as indicated by higher correlation coefficient value. Hence, the adsorption of CHCl_3 onto NZVI/ZSM can consider as a multilayer adsorption rather than a monolayer adsorption and the prepared adsorbents were found to be effect in the removal of CHCl_3 from aqueous solution.

Keywords: Water treatment; Precipitation; As(V); Arsenic removal; Sorption; Manganese

Introduction

It is well known that trace amounts of chloroform are contained in tap water and ground water. Production of chloroform in tap water is caused by reaction between residual chlorine and organic compounds [1]. Also, chloroform has been used in a wide range of industrial processes and to produce industrial products, for example, lubricants, cleaning solvents, paper bleaching, intermediates for pharmaceuticals, herbicides and fungicides [2]. Unfortunately, emissions of this compound are harmful to the environment, and in particular it is an important contributor to the destruction of the ozone layer [3]. Also it is toxic or carcinogenic and thus represents a direct health risk such as liver and kidney cancer, nervous system and reproductive effects [4-6]. Several treatment methods have been proposed for the removal of hydrocarbons especially the Low molecular weight hydrocarbons as chloroform.

Previously, the treatment of hydrocarbons on activated carbon, oxidation [7], coagulation– flocculation [8] Ion chromatography [9], Nano Chromatography and Capillary Electrophoresis [10], stripping [11] and biological treatment [12] were investigated. The removal can be performed with adsorption method using different adsorbents with high adsorption efficiency such as Nano adsorbents [13-23]. Zeolite is a type of crystalline material, containing silicon, aluminum and oxygen atoms. Zeolite can be used as catalyst as well as adsorbent material because of its high surface area and high chemical stability. By reducing the particle size, the diffusion path length decreases and hence active sites are readily accessible [22,23]. In chemistry, ZSM-5 is used to separate

molecules [24]. Its well- defined pore structure and adjustable acidity make them attractive in numerous reactions [25].

Utilization of zero-valent iron (e.g. NZVI) in the remediation application was started in the last decade [26-41]. Previously the researchers presented rapid de-halogenation of Carbon Tetrachloride (CT) and chloroform using iron particles. The aim of this study was the preparation and the characterization of the new modified adsorbents as (NZVI/ZSM) in the removal of chloroform from aqueous solution by the adsorption technique. The analysis of equilibrium data (adsorption) by model fittings is an important step to evaluate the suitable model for describing the adsorption process and the effected parameters such as contact time, initial concentration, temperature and the pH effect. Two isotherm models as Langmuir and Freundlich were used to describe the equilibrium adsorption data for chloroform adsorption. On the other hand, the kinetic data was analyzed via the pseudo-first-order and pseudo-second – order models.

Materials and Methodology

Materials

Chloroform (CHCl₃) was obtained from sigma Aldrich, Sodium borohydride (NaBH₄) of purity 98.5% and iron (III) chloride hex hydrate (FeCl₃·6H₂O) of purity 99.0% are obtained from Merck. Zeolite (ZSM-5) is purchased from Merck and the ZSM-5 with ratio of silica/Alumina (Si/Al =30) is supplied by Zeolyst International. Sodium hydroxide (NaOH) and hydrochloric acid (HCl) are obtained from Sigma-Aldrich, while analytical grade absolute ethanol and acetone are obtained from Merck and used directly without purification.

Adsorbent preparation and characterization

The ZSM-5, the commercial sample is subjected to a calcination treatment performed at 300°C for 1 h under static air environment [41], then, the impregnation of nano zero valent iron was carried out by the reduction method as sodium boro hydride reducing agent. On the other hand, the characterization of the prepared adsorbents were determined via the Fourier Transform Infra-Red (FTIR), N₂- Physisorption was used to evaluate the porosity and the surface area and the Bruner–Emmett–Teller (BET) equation was employed to calculate the specific surface area. The X-ray diffraction (XRD) and the Field Emission Scanning Electron Microscopy (FESEM).

The adsorption studies and the related parameters

The adsorptions of the Chloroform (CHCl₃) onto the adsorbents are conducted in batch mode. The required amount of adsorbent is added into the (CHCl₃) solution (of certain concentration) in the 250 mL conical flasks. The flasks are placed on the orbital shaker operating at 150 rpm in room temperature (27 ± 2°C). The initial and final concentrations of CHCl₃ are determined by using UV spectrometer (Genesys 10S

UV-VIS spectrophotometer). Detailed analysis by using UV spectrometer.

Contact time

Equilibrium time of adsorption was determined by using 0.1 g of adsorbent in 100 mL of 20 mg/L of the Chloroform (CHCl₃) solutions. The mixture was equilibrated by shaking thoroughly at 150 rpm in orbital shaker. Samples are taken out at different time intervals viz. 2, 4, 8, 15, 30, 60, 90, 120 and 180 min at pH value of 7, in room temperature (27 ± 2°C). At the end of the shaking period, the solution was sampled by a syringe. The samples were then centrifuged and the supplements are kept for further analysis. The equilibrium time was an important factor to set the adsorption time for each adsorbent.

Initial concentration

The effect of initial concentration of the (CHCl₃) on the adsorption performance is carried out through as a series of adsorption experiments with different initial concentrations ranging from 10 to 40 mg/L. The experiments are performed by using 100 mL from the (CHCl₃) solution. The adsorbent dosage is 0.1 g/L. The samples used to determine the adsorption performance are available after the batch mixture is stirred for a suitable equilibrium time (60 or 90 min).

Temperature

To study the temperature effect, the experiment is operating and the experimental were carried out in temperature range from 30-50°C.

pH

Different pH values are considered: 2, 4, 7 and 9. The pH of the solution is measured by the pH meter. The pH-meter is calibrated by using a pH buffer of 4.0, 7.0, and 10 to ensure the accuracy of the pH measurement. The initial pH of the solution is adjusted by using 1M HCl or 1M NaOH. Then, the adsorption performance is performed.

Adsorption isotherm

The isotherm studies are conducted by series of batch adsorptions. The initial pollutant concentration is ranging from 10 to 40 mg/L. Two models that are commonly used to simulate the adsorption isotherms are Langmuir and Freundlich isotherms models. The well-known expression of the Langmuir model is given in Equation 1.1:

$$q_e = \frac{K_L q_{\max} C_e}{1 + K_L C_e} \quad (1.1)$$

Where q_e is the equilibrium pollutant concentration on the adsorbent (mg/g), C_e is the equilibrium pollutant concentration in the solution (mg/L), q_{\max} is the monolayer capacity of the adsorbent (mg/g) and K_L is the Langmuir adsorption constant (L/mg) which is related to the adsorption energy. On the other hand, the Freundlich Equation can be written as Equation 1.2:

$$q_e = K_F C_e^{1/n} \quad (1.2)$$

Where q_e is the equilibrium pollutant concentration on adsorbent (mg/g), C_e is the equilibrium pollutants concentration in solution (mg/L) and K_F (mg/L)^{1/n} (g/mg)^{1/n} and n are the Freundlich constants characteristic of the system, indicators of adsorption capacity and adsorption intensity, respectively.

Kinetic studies of CHCl₃ adsorption

The kinetic studies are conducted by series of batch adsorptions running at initial pollutant concentration of 20 mg/L. The procedures of kinetic adsorption tests are identical to those of batch equilibration tests. However, the aqueous samples are taken at various time intervals. The first order rate expression based on adsorbent capacity is generally expressed as in Equation 1.3:

$$\frac{dq}{dt} = k_1(q_e - q) \quad (1.3)$$

Integration of Equation 3.5 with the boundary conditions: $t=0, q=0$, and at $t=t, q=q$, gives Equation 1.4:

$$L_n(q_e - q) = Lnq_e - k_1 t \quad (1.4)$$

Pseudo second order model is derived on the basis of the sorption capacity of the adsorbent phase, expressed as in Equation 1.5:

$$\frac{dq}{dt} = k_2(q_e - q)^2 \quad (1.5)$$

Integration of Equation 3.8 with the boundary conditions $t = 0, q = 0$, and at $t = t, q =$

q , resulted gives Equation 1.6:

$$\frac{1}{q_e - q} = \frac{1}{q_e} + k_2 t \quad (1.6)$$

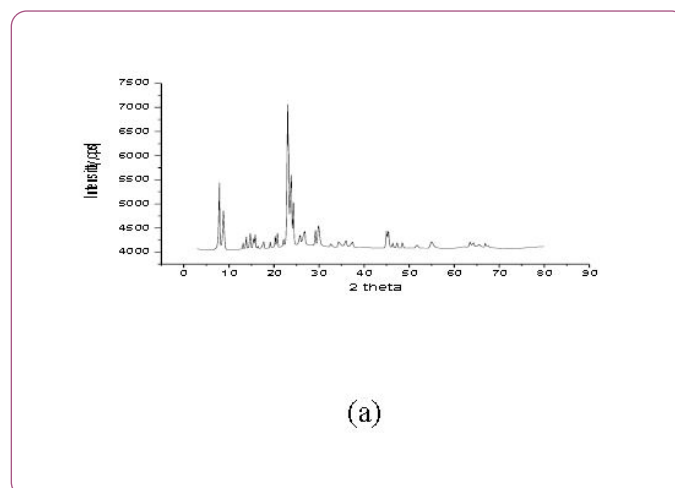
The linear form of Equation 3.8 can be written as Equation 1.7:

$$\frac{t}{q} = \frac{t}{q_e} + \frac{1}{k_2 q_e} \quad (1.7)$$

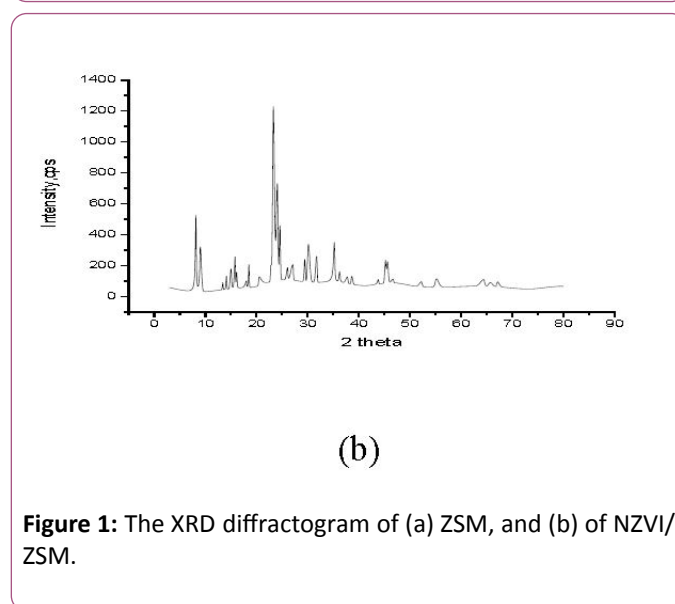
Result and Discussion

The adsorbent characterization The XRD analysis

The XRD pattern of the ZSM-5 is illustrated in **Figure 1a**. The main peaks appear at 2θ of $7.92^\circ, 8.76^\circ, 15.84^\circ, 23.6^\circ, 37.48^\circ, 45.39^\circ$ and 54.94° in the detection of (1 0 1), (2 0 0), (2 0 2), (3 3 2), (2 8 0), (0 8 4) and (1 0 8), respectively, indicating the crystallization of ZSM-5. The XRD diffractions of NZVI/ZSM were illustrated in **Figure 1b**. The presence of Zero Valent Iron is indicated by a distinct peak at 2θ of 44.45° with detection (1 1 1), which is related to -Fe (NZVI) as observed in the diffractogram of NZVI/ZSM. It appears that the ZeroValent Iron is successfully loaded into the supports.



(a)



(b)

Figure 1: The XRD diffractogram of (a) ZSM, and (b) of NZVI/ZSM.

N₂-Physisorption analysis of adsorbents nano zero valent iron (NZVI)

Parameters such as BET surface area pore volume and pore size of adsorbents are summarized in **Table 1**.

Table 1: BET surface area (SBET), pore volume and pore size of the prepared adsorbents.

Adsorbent	BET m ² /g	BJH adsorption Pore volume cm ³ /g	BJH adsorption Pore size Å
NZVI	40.65	0.165	152.70
ZSM	254.33	0.095	63.28
NZVI/ZSM	116.52	0.055	91.22

The surface areas of NZVI are found to be 40.65 m²/g. Meanwhile, their pore sizes is 152.70Å. It is clear that the surface area and the pore size of NZVI obtained by this study is larger than that reported previously [42], which may be attributed to different experimental conditions. Therefore, the surface area of the adsorbent is dependent on the particle

size, morphology, surface texturing and porosity of the prepared adsorbent. On the other hand, the N₂ adsorption-desorption isotherms for unsupported adsorbents (type I isotherm) depend on (IUPAC) classification at 77K illustrated in **Figure 2**.

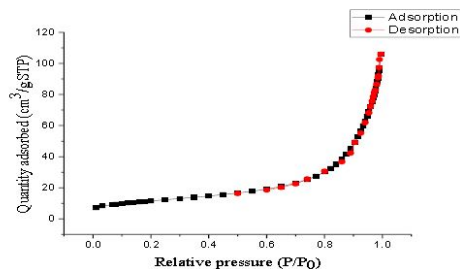


Figure 2: N₂ adsorption/desorption equilibrium isotherms at 77 K for NZVI.

FTIR Analysis

Nano zero valent iron (nzvi)

Figure 3 shows broad peaks at around 3200–3600 cm⁻¹ in both spectra. These peaks are related to the OH vibration stretching band as presented in **Figure 3** clearly. These -OH

groups may come from the moisture water. The peak at a wavenumber of 1653 cm⁻¹ observed may correspond to the bending modes of OH group [43]. As shown in peak at 1427 cm⁻¹ is related to the Fe-O stretching vibration, peak at 1328 cm⁻¹ is related to the Fe-O bending out of the structure and peak at 710 cm⁻¹ is related to is attributed to the NZVI (Fe₀) as reported by the previous studies [44,45].

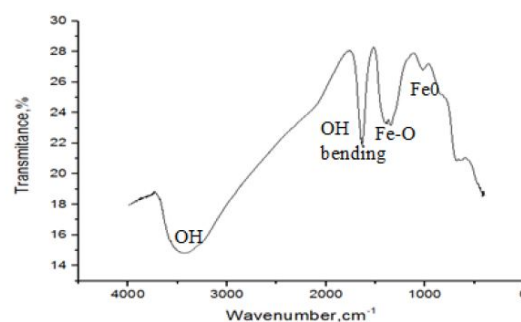


Figure 3: FTIR spectrum of NZVI.

The FTIR spectra of NZVI/ZSM are shown in **Figure 4**. The broad spectrum of adsorption bands at wavenumbers of 3800 cm⁻¹ and 3400 cm⁻¹ are related to the intermolecular OH (Si-OH-Si and Al-OH- cm⁻¹ [46]. Also, they are related to H-O-H stretching. The band at wavenumbers of 1648 cm⁻¹ and 1661 cm⁻¹ can be attributed to O-H bending [47].

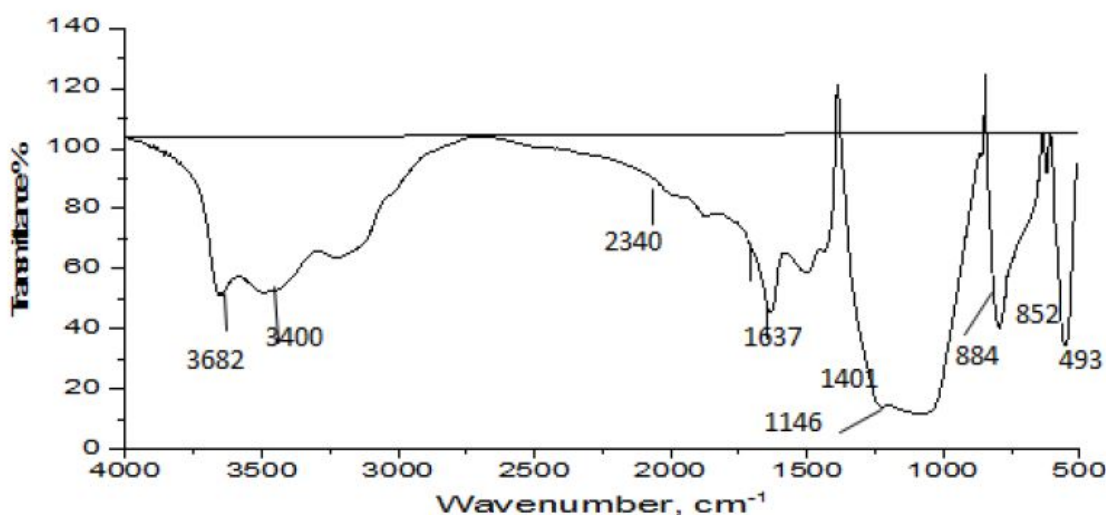


Figure 4: FTIR spectrum of ZSM-5.

FESEM-EDX for ZSM and NZVI/ZSM

The SEM micrograph for ZSM-5 is illustrated in **Figure 5** showing the surface morphology of ZSM-5. Irregular crystallite

particles are observed, similar to those reported by others [48,49].

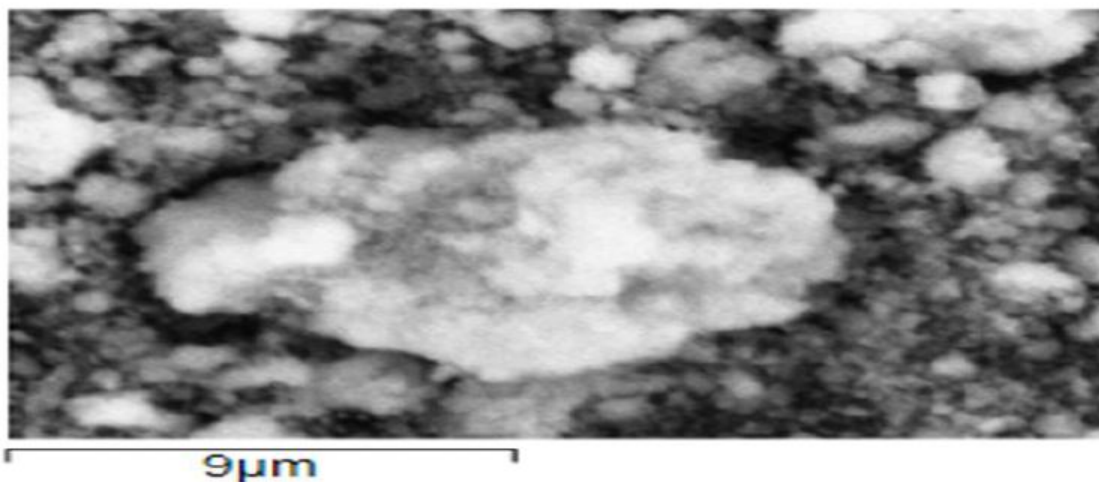


Figure 5: SEM micrograph for ZSM-5.

The EDX spectra, **Figure 6** for ZSM-5 appears that the peak of the value 1.8 KeV was related to the Si, while peak in the value of 1.6KeV was related to Al and peak in the value of 0.6 KeV, was related to the O atom.

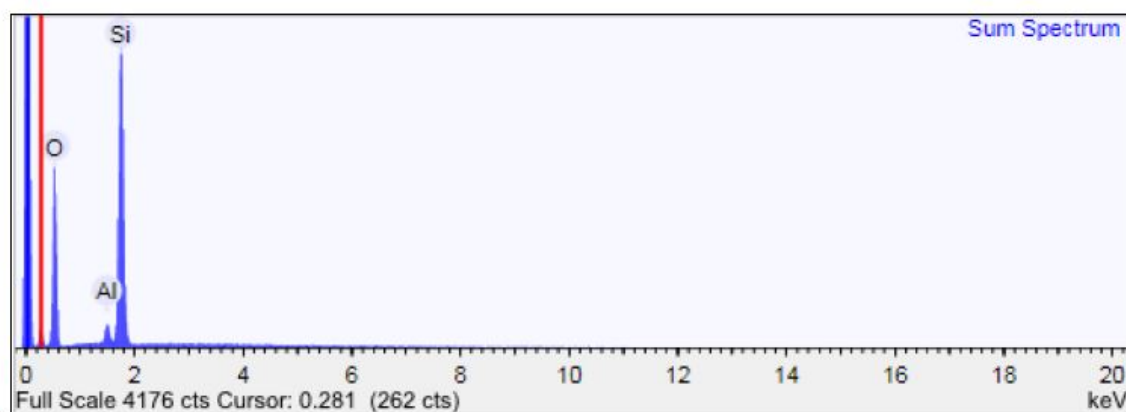


Figure 6: The EDX spectra of ZSM-5

Table 2: The EDX analysis of ZSM, and NZVI/ZSM.

Adsorbents	Element%	Weight%	Atomic weight %
ZSM-5	O	60.86	73.15
	Si	37.21	25.47
	Al	1.93	1.38
NZVI/ZSM	O	75.63	85.69
	Si	18.55	11.97
	Al	1.29	0.87
	Fe	4.53	1.47

The EDX analysis as shown in **Table 2** appears the weight and the atomic weight for the elements in ZSM-5 of the entity of silicon, aluminium and oxygen percentage. As appear in **Table 2**, the weight and the atomic weight for silicon,

aluminum and oxygen percentage in the ZSM-5 surface. The percentage of Si atoms in the ZSM-5 at the atomic value of 25.47% with the weight value of 37.20%, the O atoms in the ZSM-5 at the atomic value of 73.15% with the weight value of 60.86% and the Al atoms in the ZSM-5 at the atomic value of 1.37% with the weight value of 1.93%, respectively.

The mapping distributions (from EDX analysis) for oxygen, aluminium, silicon and mixture of the three elements on the

surface of ZSM-5 are presented in **Figure 7a-d** respectively. It is obvious that the oxygen distribution on the surface is homogeneous as shown in **Figure 7a**. But, the amount of aluminium (see **Figure 7b**) seems to be less in the structure of ZSM-5 as compared to that of Si as shown in **Figure 7c**. On the other hand, **Figure 7d** shows the distribution for the three elements on the surface of ZSM-5.

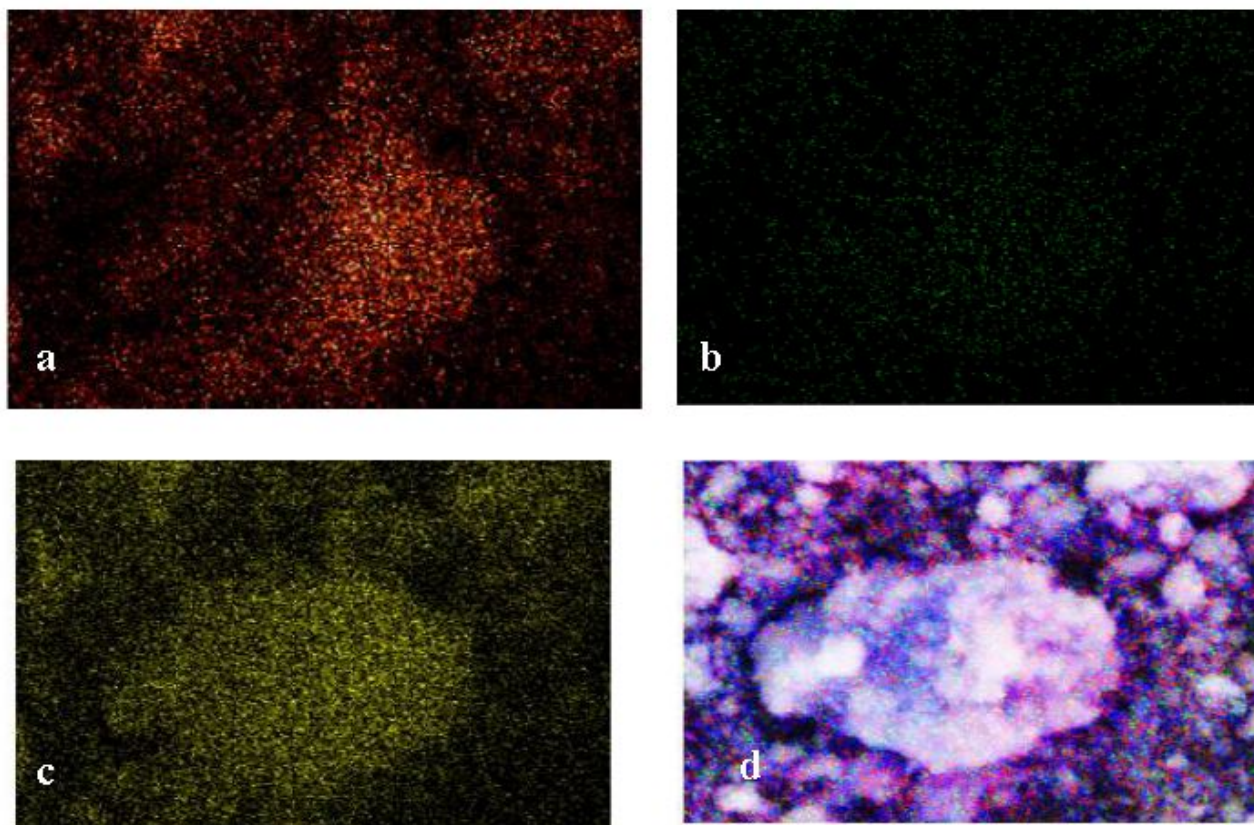


Figure 7: EDX mapping distribution for (a) oxygen, (b) aluminum, (c) silicon and (d) mixture for three elements for ZSM-5 surface.

Adsorption study

The adsorption properties of NZVI/ZSM were evaluated in sorption of CHCl_3 on the other hand, CHCl_3 consider as low solubility in water, C_s of Chloroform is 68.67 mmole/dm^3 of Chloroform Thus, CHCl_3 stock solution prepared via the dissolving in alcohol as 2- propanol instead of distilled water.

Adsorption isotherm of CHCl_3 onto nzvi/zsm

Adsorption isotherm studies are performed by batch adsorption of CHCl_3 at various initial concentrations ranging from 10-40 mg/L. Triplicate the data collection to evaluate the

vertical error bars and the average deviation on the data points. The Langmuir and Freundlich constants of adsorption of CHCl_3 onto NZVI/ZSM are 0.0368 L/mg and 1.1123 g/L respectively. The maximum adsorption capacities of CHCl_3 onto NZVI/ZSM and determined by the Langmuir isotherm are 19.92 mg/g. In general, Freundlich isotherm can better characterize the adsorption process for NZVI/ZSM. The surface heterogeneities of NZVI/ZSM are prevalent because $1/n \sim 1.0$ [50]. Hence, the adsorption of CHCl_3 onto NZVI/ZSM can be considered as multilayer adsorption. The Langmuir and Freundlich isotherm models for adsorption of CHCl_3 onto NZVI/ZSM was presented in **Figure 8a** and **8b** respectively.

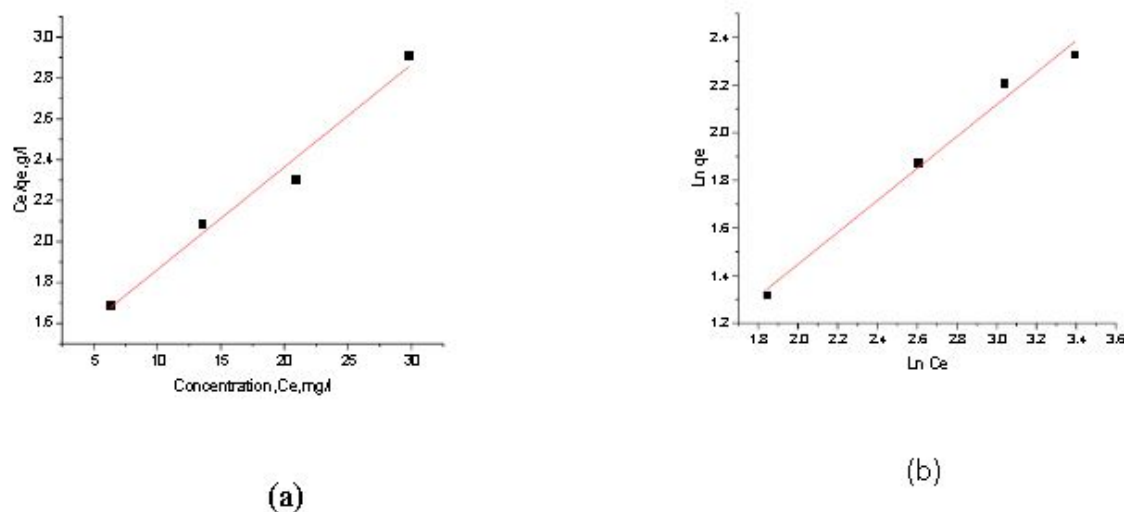


Figure 8: Langmuir (a) and Freundlich (b) adsorption isotherm model for sorption of CHCl₃ onto NZVI/ZSM.

The parameters corresponding to the fitting of these results to the Langmuir and Freundlich isotherm models are summarized in **Table 3**. The maximum adsorption uptakes values of the prepared adsorbents are comparable with some other adsorbent as shown in the **Table 3**.

Table 3: Isotherm parameters of CHCl₃ adsorption onto NZVI/ZSM with comparison by uptake value from the literatures.

Absorbents	Langmuir			Freundlich			Reference
	KL L/mg	qmax, mg/g	R2	KF	1/N	R2	
NZVI/ZSM	0.0368	19.92	0.9786	1.1123	0.67	0.9872	This study
Activated carbon SKD515	5.47	9.69	0.9960	12.15	0.78	0.9500	[4]
Nano-TiO ₂	-	57.6	-	-	-	-	[3]
Rectorite/ Chitosan, nanocomposite	0.7	6.3	-	2.3	0.7	-	[9]

Kinetic studies of CHCl₃ onto NZVI/ZSM

This experiment was carried out at the variation of time of contact (0 to 180 min). Generally, adsorption increases during the first hour and it decays thereafter. The kinetic experiments are similar to the batch equilibrium tests; however, the samples are taken at certain time instant within 180 min. The kinetic studies are conducted by a series of batch adsorptions at initial CHCl₃ concentration of 20 mg/L. Linear plot of both

first and second order kinetic models are illustrated in **Figure 9a and b** while the kinetic parameters calculated were summarized in **Table 4**. Linear regression of both models provided the correlation coefficient value of 0.9439 for pseudo-first order and value of 0.9944 for pseudo second order for NZVI/ZSM. In **Table 4** the q_e experimental data of pseudo-second order near to the calculate value of the q_e Cal by the slope of this model. As result, the pseudo second order model is more suitable to be applied in this study (higher R²).

Table 4: Kinetic parameters of the adsorption of CHCl₃ onto NZVI/ZSM were collected in this study

Adsorbent	k ¹ , min ⁻¹	q _e , Cal mg/g	R ²	k ² , g/mg min	q _e , mg/g	Cal	R ²	q _e mg/g
NZVI/ZSM	0.030	7.75	0.9439	0.11	12.53		0.9944	12.08

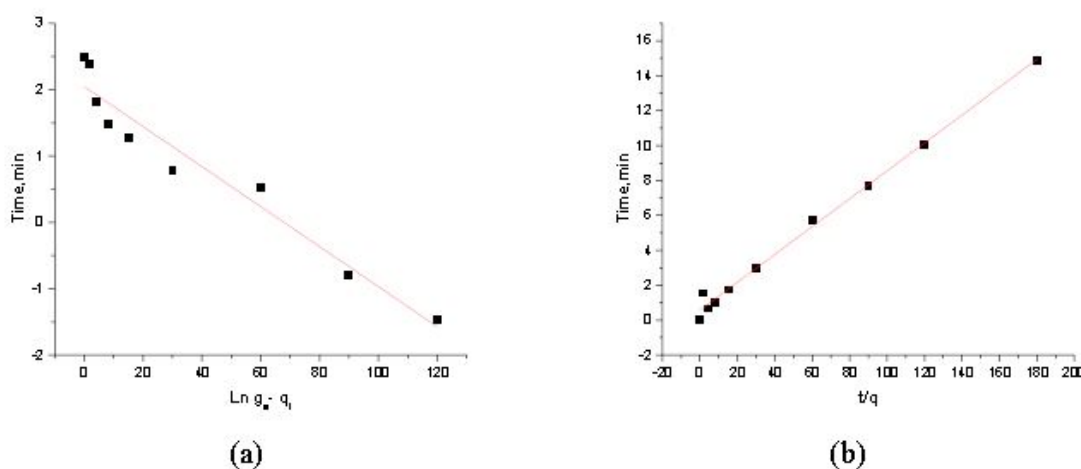


Figure 9: Liner plot of pseudo-first-order (a) and pseudo-second-order (b) kinetic models for adsorption of CHCl₃ onto NZVI/ZSM.

Effect of temperature

The effects of temperature on the adsorption equilibrium of CHCl₃ onto NZVI/ZSM was presented in **Figure 10**. Generally, the adsorption of CHCl₃ increases as the temperature increases from 30°C to 50°C. As result, the uptake of NZVI/ZSM with a value of 15.499 mg/g and removal of 78.56% at 50°C.

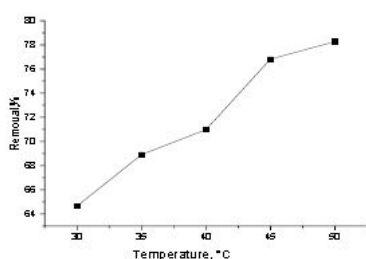


Figure 10: Effect of adsorption temperature on CHCl₃ adsorption onto NZVI/ZSM removal.

Effect of pH of CHCl₃ adsorption on to NZVI/ZSM

Affected factor on the adsorption of CHCl₃ is the pH of the solution. It is generally expected that adsorption increases as pH increases. As shown in **Figure 4**, the removal of CHCl₃ increases significantly as pH increases from 2 to 9. In this study, the removal rate of CHCl₃ is the highest as pH=9. Initial pH of the solution was adjusted by using 1M HCl or 1M NaOH.

The effects of the pH solution of the adsorptions of CHCl₃ onto NZVI/ZSM is shown in **Figure 11**. Here, the removal of CHCl₃ increases significantly as pH increases from 2 to 9. In this study, the highest CHCl₃ removal is attained when pH=9 with

CHCl₃ removal as high of 50.37% for NZVI/ZSM at CHCl₃ initial concentration of 20 mg/L.

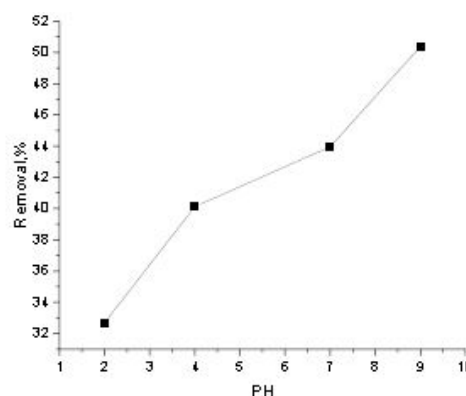


Figure 11: Effect of adsorption pH onto CHCl₃ adsorption onto NZVI/ZSM removal.

Conclusion

In this study, the preparation and characterization of NZVI/ZSM was determined via the adsorption studies in batch mode. The evaluation of the adsorption performance of the prepared adsorbents were proven via the isotherm studies by Langmuir and Freundlich models, Freundlich isotherm can better characterize the adsorption process of CHCl₃ onto NZVI/ZSM. Hence, the adsorption of CHCl₃ onto NZVI/ZSM can be considered as multilayer adsorption. On the other hand, the kinetic studies by pseudo first and second order model were carried out for the adsorption of CHCl₃ onto NZVI/ZSM. The pseudo first order and the pseudo second order kinetic models are examined. It seems that the pseudo second order kinetic model fits well with the experimental data. The effect of the

adsorption efficiency of NZVI/ZSM was increased with increased temperature and pH to 9. From the all above, the ZSM-5 as supporting materials is perfect in stabilizing the NZVI; increase the adsorption efficiency at value of 19.92 mg/g and this adsorbent is comparable with other adsorbents from the literature. These new adsorbents prevent the agglomeration of the Nano zero valent iron (NZVI) and could be employed as efficient applicable and suitable adsorbents for the removal of CHCl_3 from aqueous solutions.

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References

1. Background document for development of WHO Guidelines for Drinking-water Quality, World Health Organization 2004.
2. Chen ChH, Dural NH (2002) Chloroform adsorption on soil. *Chem Eng J* 47: 110-1115.
3. Gharbani P, Mehrizad A, Ataie S, Mosenharzandi A, Tabatabaei SM, et al. (2011) Adsorption of chloroform from aqueous solution by nano - TiO_2 . *Int J Nano Dimens* 1: 287-296.
4. Hamdaouia O, Naffrechoux E (2007) Modeling of adsorption isotherms of phenol and chlorophenols onto granular activated carbon. Part I. Two-parameter models and equations allowing determination of thermodynamic parameters. *J Hazard Mater* 147: 381-394.
5. Insa S, Salvado V, Antico E (2006) Assays on the simultaneous determination and elimination of chloroanisoles and chlorophenols from contaminated cork samples. *J Chromatogr A* 1122: 215-221.
6. Kadirova ZC, Okada K, Katsumata K, Isobe T, Matsushita N, et al. (2013) Adsorption and photodegradation of methylene blue by iron oxide impregnated on granular activated carbons in an oxalate solution. *J Appl Surf Sci* 284: 72-79.
7. Krishnaiah D, Sarbatly R, Anisuzzaman SM, Bono A (2013) Adsorption of 2,4,6- Trichlorophenol (TCP) onto activated carbon. *J King Saud University* 25: 251-255.
8. Liu X, Li X, Yang Q, Yue X, Shern T, et al. (2012) Landfill leachate pretreatment by coagulation-flocculation process using iron-based coagulants: Optimization by response surface methodology. *Chem Eng J* 200: 39-51.
9. Michal ski R (2006) Instrumental methods in metal ions speciation: Chromatography, Capillary Electrophoresis and Electrochemistry. *Crit Rev Anal Chem* 36: 107-127.
10. Ali I, Aboul-Enein H, Gupta V (2009) Nano Chromatography and Capillary Electrophoresis: Pharmaceutical and Environmental Analyses. Wiley & Sons, Hoboken, USA.
11. Srinivasan A, Chowdhury P, Viraraghavan T (2008) Air Stripping in Industrial Wastewater Treatment. *Encyclopedia of Life Support Systems (EOLSS)*.
12. Robinson T, Mcmultan G, Marchant R, Nigam P (2001) Remediation of dyes in textile effluent: a critical review on current treatment technologies with a proposed alternative. *Bio Resource Technology* 77: 247-255.
13. Ali I (2012) New Generation Adsorbents for Water Treatment. ACS publication 112: 5073-5091.
14. Ali I (2014) Water Treatment by Adsorption Columns: Evaluation at Ground Level. *Separation and purification reviews* 43: 175-205.
15. Ali I, Gupta V (2006) Advances in water treatment by adsorption technology. *Nature Protocols* 1: 2661-2667.
16. Ali I, Aboul Enein H (2002) Speciation of arsenic and chromium metal ions by reversed phase high performance liquid chromatography. *Chemosphere* 48: 275-278.
17. Ali I, Gupta V, Khan T, Asim M (2012) Removal of Arsenate from Aqueous Solution by Electro-Coagulation Method Using Al-Fe Electrodes. *Int J Electrochem Sci* 7: 1898-1907.
18. Meher A, Das S, Rayalu S, Bansiwala A (2016) Enhanced arsenic removal from drinking water by iron-enriched aluminosilicate adsorbent prepared from fly ash. *Desal Wat Treat* 57: 44.
19. Ali I, Khan T, Asim M (2012) Removal of arsenate from groundwater by electrocoagulation method. *Environ Sci Pollut Res* 19: 1668-1676.
20. Ali I, AL-Othman Z, Alwarthan A (2016) Molecular uptake of congo red dye from water on iron composite nano particles. *J Mol Liq* 224: 171-176.
21. Ali I, AL-Othman Z, Alwarthan Sanagi M (2015) Green Synthesis of Iron Nano-Impregnated Adsorbent for Fast Removal of Fluoride from Water. *J Mol Liq* 211: 457-465.
22. Alharbi O, Basheer A, Khahab R, Ali I (2018) Health and environmental effects of persistent organic pollutants. *J Mol Liq* 263: 442-453.
23. Ali I, Alharbi O, AL-Othman Z, Alwarthan A (2018) Facile and eco-friendly synthesis of functionalized iron nanoparticles for cyanazine removal in water. *Colloids Surf B Biointerfaces* 171: 606-613.
24. Kirsanov MP, Shishkin VV (2016) Evaluation and improving the efficiency of the use of activated carbon for the extraction of organic chlorine compound in water technology. *J Food Raw Mater* 4: 148-153.
25. Li Sh, Ding L, Zhou P (2011) Adsorption application for removal of hazardous Chloroform from aqueous solution by Nano composites Recto rite/Chitosan adsorbent. *J Water Resource Prot* 3: 448-455.
26. Ling X, Li J, Zhu W, Zhao Y, Sheen J, et al. (2012) Synthesis of Nano scale zero-valent iron/ordered mesoporous carbon for adsorption and synergistic reduction of nitrobenzene. *J Chemosphere* 87: 655-660.
27. Ali I, Alharbi O, Allothman Z, Badjah A, Alwarthan A, et al. (2018) Artificial neural network modeling of amid black dye sorption on iron composite Nano material: Kinetics and thermodynamics studies. *J Mol Liq* 250: 1-5.
28. Burakova E, Dyachkova T, Rudov A, Tugolukov E, Glalun E, et al. (2018) Novel and economic method of carbon nanotubes synthesis on a nickel magnesium oxide catalyst using microwave radiation. *J Mol Liq* 253: 340-346.
29. Ali I, Khan T, Asim M (2011) Removal of Arsenic from Water by Electrocoagulation and Electro dialysis Techniques. *Separation and purification reviews* 40: 1.
30. Ali I, Allothman Z, Alwarthan A (2017) Supra molecular mechanism of the removal of 17- β - estradiol endocrine

- disturbing pollutant from water on functionalized iron Nano particles. *J Mol Liq* 241: 123-129.
31. Ali I, Alothman Z, Alwarthan A (2017) Uptake of propranolol on ionic liquid iron Nano composite adsorbent: Kinetic, thermodynamics and mechanism of adsorption. *J Mol Liq* 236: 205-213.
 32. Ali I, Alothman Z, Alwarthan A (2016) Molecular uptake of congo red dye from water on iron composite nano particles. *J Mol Liq* 224: 171-176.
 33. Dehghan M, Sanaei D, Ali I, Bhatnagar A (2016) Removal of chromium(VI) from aqueous solution using treated waste newspaper as a low-cost adsorbent: Kinetic modeling and isotherm studies. *J Mol Liq* 215: 671-679.
 34. Ali I, Jain Ch (2004) Advances in arsenic speciation techniques. *Int J Anal Chem* 8: 12.
 35. Suhail M, Ali I (2017) Advanced spiral periodic classification of the Elements. Poster.
 36. Ali I, Alothman Z, Alwarthan A (2016) Sorption, kinetics and thermodynamics studies of atrazine herbicide removal from water using iron Nano-composite material. *Inter J Environ Sci Technol* 13: 733-742.
 37. Ali I, Alothman Z, Alwarthan A (2016) Removal of sebumeton herbicide from water on composite Nano adsorbent. *J Desalin Water Treat* 57: 10409-10421.
 38. Ali I, Alothman Z, Alwarthan A (2016) Uptake of pantoprazole drug residue from water using novel synthesized composite iron nano adsorbent. *J Mol Liq* 218: 465- 472.
 39. Ali I, Alothman Z, Alwarthan A (2016) Green synthesis of functionalized iron Nano particles and molecular liquid phase adsorption of ametryn from water. *J Mol Liq* 221: 1168-1174.
 40. Ali I, Alothman Z, Alwarthan A (2016) Synthesis of composite iron nano adsorbent and removal of ibuprofen drug residue from water. *J Mol Liq* 219: 858-864.
 41. Ali I, Asim M, Khan T (2013) Arsenite removal from water by electro-coagulation on zinc– zinc and copper–copper electrodes. *Inter J Environ Sci Technol* 10: 377-384.
 42. Noubactep C, Caren R, Care S (2012) Nano scale metallic iron for environmental remediation: prospects and limitations. *J Water Air Soil Pollut* 223: 1363-1382.
 43. Park J, Kim H, Han Y (2012) Formation of mesoporous materials from silica dissolved in various NaOH concentrations: Effect of pH and ionic strength. *J Nano Mater* 2012: 1-10.
 44. Petala E, Karakassides MA, Dimos K, Zboril R, Douvalis A, et al. (2013) Nano scale zero-valent iron supported on mesoporous silica: Characterization and reactivity for Cr(VI) removal from aqueous solution. *J Hazard Mater* 216: 295-306.
 45. Sun X, Wang L, Yan Y, Li J, Han W (2014) SAB-15-incorporated nanoscale zero- valent iron particles for chromium(VI) removal from groundwater: Mechanism, effect of pH, humic acid and sustained reactivity. *J Hazard Mater* 12: 26-33.
 46. Tahir NM, Ariffin MM, Khoon TS, Hui TJ, Suratman S (2008) Adsorption of chloroform –ethyl and metasulfuron-methyl on selected Selangor agricultural soils. *T Mal J Anal Sci* 12: 341-347.
 47. Tay T, Ceylan B, Erdem M, Karagoz S (2012) Adsorption of methylene blue from aqueous solution on activated carbon produced from soybean oil cake by KOH activation. *J Bioresour* 7: 3175-3187.
 48. Tan IW, Hameed BH, Ahmad L (2008) Adsorption of basic dye on high-surface-area activated carbon prepared from coconut husk: Equilibrium, kinetic and thermodynamic studies. *J Hazard Mater* 154: 337-346.
 49. Tebeje Z (2011) GAC Adsorption processes for chloroform removal from drinking water. *Ta J N Appl Sci* 21: 352-358.
 50. Wang P, Cao M, Qian J, Wang Ch, Ao Y, et al. (2014) Kinetics and thermodynamics of adsorption of methylene blue by a magnetic graphene-carbon nanotube composite. *J Appl Surf Sci* 290: 116-124.